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Biomonitoring of atmospheric particulate pollution via chemical composition and magnetic properties of roadside tree leaves

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Abstract

Particulate matter (PM) is a main atmospheric pollution which threatens human health and well-being. In this research, we chemically and magnetically analysed roadside tree leaves, collected from three tree species in two main roads (from two different cities) and a reference area, for 28 elements and the saturation isothermal remanent magnetisation. Comparison of unwashed and washed leaves revealed that deposited particles on the leaf surface contain various elements including Al, Ca, Fe, Mg, Mn, Na, Si, Ti, Ba, Co, Cr, Cu, Ni, Rb, V, Zn, and Zr. Moreover, there was no significant difference between washed/unwashed leaves in Cl, K, P, S, As, Cd, Cs, Pb, Sn and Sr concentrations, which indicates tree leaves may not be a suitable biomonitor for these elements. Our results showed that site and tree species are important factors which affect atmospheric

30 elements deposition. Among the three considered tree species, *Chamaecyparis*
31 *lawsoniana* showed the highest potential for atmospheric particle accumulation.
32 The PCA results revealed that Al, Fe, Ti, Co, Cr, Cu, Ni, Rb, Si, V, Zn, and Zr
33 indicated emissions from road traffic activities and soil dust, and Ca, Mg and Na
34 from sea salts, and Mn and Sb from industrial activity. The biplot results showed
35 that the site effect was much stronger than the species effect for all elements and
36 SIRM values. Moreover, elements from traffic, industrial activity and soil dust are
37 significantly correlated with leaf SIRM indicating that leaf SIRM can be a suitable
38 bioindicator of exposure to traffic-derived particles and soil dust, and not from
39 sea salts. It is concluded that chemical composition and SIRM of urban tree
40 leaves can serve as a good indicator of atmospheric PM pollution in Iran, and
41 anywhere else where the studied trees grow.

42

43 Key words: chemical composition; atmospheric deposition; biomagnetic;
44 biomonitoring; tree leaves; species specific

45

46 **1 Introduction**

47 Particulate matter (PM) with aerodynamic diameter below 2.5 and 10 μm (PM_{2.5}
48 and PM₁₀) are pollutants which can seriously harm human health and wellbeing in
49 urbanized and industrialized areas. The main sources of urban atmospheric
50 particulates are road activity, industrial activity, domestic heating and
51 construction activity. Among various sources of PM, road traffic induced particles
52 are known as a main source of suspended particles in many urban atmospheres
53 (Vu et al. 2015); they include exhausted emissions, wear and tear fractions, and
54 resuspended roadside dust due to vehicles movements and wind. Chemical
55 composition of traffic-derived particles contain a number of trace metals including
56 Fe, Mn, Cu, Ni, Zn, Ba, Cd, Pb, Cr, Ti and V (Huhn et al. 1995; Harrison and Jones
57 1995; Gunawardena et al. 2012) of which some have magnetic properties (Maher
58 et al. 2008). Metals can be inhaled by humans and cause inflammatory reactions

59 or may adhere to suspended atmospheric particles and deposit on different
60 surfaces. These metals are mainly toxic and persistent, and can be taken up by
61 different plants and animals and transferred through the food-chain to humans.
62 Many researchers formerly used biological material for monitoring of air quality,
63 in particular PM pollution (Conti and Cecchetti 2001; Kardel et al. 2010; Kardel et
64 al. 2011; Vuković et al. 2015), as the measurement of PM concentration directly
65 in the atmosphere is expensive and practically impossible at high spatial
66 resolution. Among different biological material, plants - specifically trees - provide
67 some services such as increasing carbon storage, mitigating the heat-island effect
68 and wind storms, decreasing sound pollution, improving air quality in an urban
69 environment (Rai 2016) and protection of cultural heritage (Kocić et al. 2014).
70 Urban trees, with dense twigs and leaves and complex leaf surface morphology,
71 as natural receptors can effectively capture and reduce atmospheric particles
72 (Jamil et al. 2009; Fantozzi et al. 2013; Nowak et al. 2006; Wang et al. 2006).
73 Different methods can be applied for the monitoring of deposited atmospheric
74 particles on leaf surfaces. Due to the strong correlation between atmospheric PM
75 and NO_x concentrations and heavy metal concentrations on the one hand and
76 magnetic properties of leaves on the other hand (Hofman et al. 2017; Maher et
77 al. 2008), magnetic measurements of leaves demonstrated to be a good proxy for
78 estimating the atmospheric, traffic and industrial particle pollution (Hofman et al.
79 2014 and 2017; Kardel et al. 2011 and 2012; Lu et al. 2008; Maher et al. 2008;
80 Mitchell et al. 2010). Saturation isothermal remanent magnetisation (SIRM) is
81 one of the magnetic variables with high sensitivity for materials which are weak in
82 magnetic properties (Mitchell et al. 2010). Despite the large amount of studies
83 already performed on SIRM of different tree species in different urban
84 environments (Hofman et al. 2014; Kardel et al. 2011 and 2012; Maher et al.
85 2008; Mitchell et al. 2010; Sant'Ovaia et al. 2012), there is still not enough
86 information on SIRM and large number of metal concentrations of deposited
87 particles on tree species (with different leaf surface morphology) in sites with

88 different climates and traffic densities. Also, it is not fully clear which one of the
89 two, site or species, effects the accumulation of particles on the leaf's surface
90 more and which elements can differentiate one site from another the most.

91 In Iran, due to high air pollution problems in mega cities, the monitoring of air
92 quality for other cities is largely ignored. The aim of this study was (i) to
93 determine and identify the elements deposited on the leaf surface of three tree
94 species in three different road sites in different climates and traffic densities; (ii)
95 to assess the effect of species and sites on the deposited elements; and (iii) to
96 evaluate the relationship between large number of metal concentrations of leaves
97 and leaf saturation isothermal remanent magnetization (SIRM).

98

99

100 **2 Materials and methods**

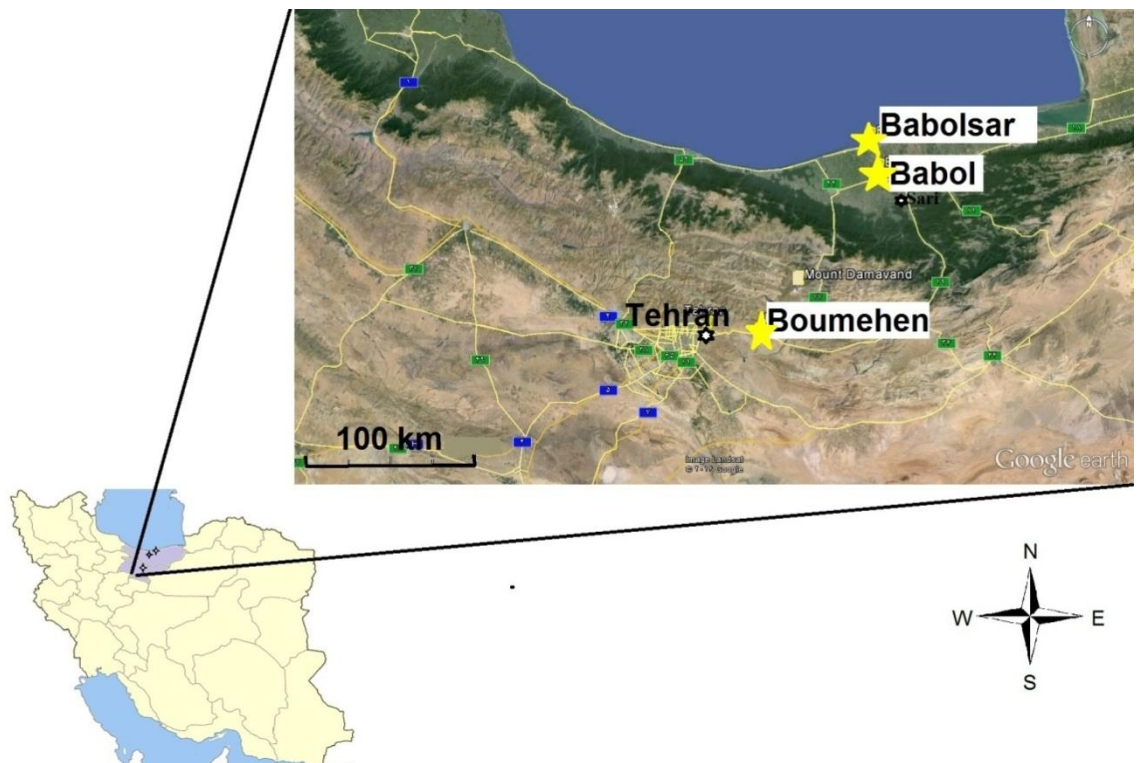
101

102 **2.1 Study area and sampling**

103 In each of the cities Babolsar (36° 32' 1.80" N, 52° 38' 56.23"E), Babol (36° 41'
104 52.22" N, 52° 35' 36.12" E) and Boumehen (35° 43' 55.40" N, 51° 52' 6.65"E)
105 in Iran (Fig. 1), we selected a study site in order to compare the chemical
106 composition of leaf deposited atmospheric particulate matter. The city of
107 Boumehen is approximately 1960 m above sea level, Babolsar 21 m, and Babol
108 city is 2 m below the sea level. Babol city is located 20 km from Babolsar city and
109 164 km from Boumehen. The prevailing wind direction in the year of sampling
110 was northeast, west and northeast, and sum of precipitation was 1110, 764,
111 631mm in the cities of Babolsar, Babol and Boumehen, respectively, as measured
112 in the nearest weather stations (data obtained from I.R. of Iran Meteorological
113 Organisation, <http://www.irimo.ir>). The mean temperature in the cool season
114 (winter time) was 8, 7 and 6.5°C, and in the hot season (summer time) it was
115 26, 26 and 30°C in the cities of Babolsar, Babol and Boumehen, respectively.

116 The road from Babol city was a boulevard with a width of around 30 m and high
117 traffic density (more than 20000 vehicles each day in this road). In the city of
118 Boumehen, at 30 km to the capital city Tehran, a road with high traffic density
119 was selected. The city has different industrial activities (e.g., cement production,
120 silica sand mining, hard rock, sand and gypsum mining, ceramic factory) at
121 different distances (25-60 km) to the city centre. A local road from the sub-urban
122 area of Babolsar city was selected as a reference location, with a distance of
123 about 100-250 m to the southern Caspian Sea. In Iran, vehicles use different
124 types of fuel, i.e. petrol, gasoline, compressed natural gas (CNG) and liquid
125 petroleum gas (LPG).

126



127
128

129 Fig. 1: study area indicating the three sampling sites (Babolsar, Babol and
130 Boumehen) in the northern part of Iran (source: Google Maps)

131

132 Three evergreen species were selected, i.e. *Chamaecyparis lawsoniana* (English
133 name: Lawson falcyl) with scale-like leaves, *Ligustrum japonicum* (English name:
134 wax-leaf privet) with broad leaves, and *Pinus brutia* subsp. *eldarica* (Medw) Silba

135 (English name: Eldar pine) with needles based on their availability near the
136 selected roads.

137 In Babolsar, *P. brutia* subsp. *eldarica* trees were located at a distance of 150-250
138 m to the beach, while the *C. lawsoniana* trees were at 100-150 m to beach.

139 *L. japonicum* was sampled only in Babol. *C. lawsoniana* was sampled from the
140 same site as *L. japonicum* in Babol and in Babolsar. *Pinus brutia* subsp. *eldarica*
141 was sampled in the cities of Babolsar and Boumehen. Leaf samples were collected
142 from 46 trees (10 trees from each species in the city of Babol and Boumehen, and
143 8 trees from each species in the city of Babolsar) at a height of 1.5-2 m from
144 ground level on 2 and 3 July 2012. There was no rain at least for 10-15 days
145 before sampling. Leaf samples from each tree were stored in a paper envelope
146 and transferred to the laboratory for chemical and magnetic analyses.

147

148 **2.2 Sample preparation and XRF analysis**

149 In the lab, the samples were divided into two subsamples: one subsample was
150 washed with 20 ml of deionised water in a glass beaker and gently rubbed for one
151 minute with the fingers covered by gloves (for each leaf separately) to remove
152 particles deposited on the leaf surface, and the other subsample remained
153 unwashed. A subsample contained 1-2 leafy branchlets of *C. lawsoniana*, 5-6
154 leaves of *L. japonicum* or 6-8 needles of *Pinus brutia* subsp. *eldarica*. By
155 comparison of the washed and unwashed subsamples, we can identify the
156 elements that belong to deposited particles on the leaf surface, not to the leaf
157 itself. Both washed and unwashed leaves were dried in an oven for 7 days at 45
158 °C. Then, the subsamples were powdered using a mortar and pestle.
159 Subsequently, the weights of the subsamples were determined by an electronic
160 balance with an accuracy of 0.0001 g.

161 For making a pellet, around 0.6 g glucose was pressed with 5 ton for 30 s. Then,
162 0.48 g sieved leaf powder was gently mixed with 0.08 g glucose and added to the
163 glucose pellet surface. Thereafter, these double layer samples were pressed with

164 8 ton for 30 s. Samples were covered with microfilm and analyzed using energy
165 dispersive X-ray Fluorescence (PANalytical, Epsilon 5) for the concentration of 28
166 elements (Na, Mg, Al, Si, P, S, Cl, K, Ca, Ti, V, Cr, Mn, Fe, Co, Ni, Cu, Zn, As, Rb,
167 Sr, Zr, Cd, Sn, Sb, Cs, Ba and Pb).

168

169 **2.3 Sample preparation and SIRM analysis**

170 Between 4 to 8 leaves, branchlets or needles were tightly packed by cling film
171 into a 10cc plastic container for magnetic measurement. The containers with the
172 samples were magnetized with a pulsed DC magnetic field of 1 Tesla using a
173 Molspin pulse magnetiser (Molspin Ltd, UK) as described by Matzka and Maher
174 (1999). Subsequently, the saturation isothermal remanent magnetisation (SIRM)
175 was measured using a Molspin Minispin magnetometer with high sensitivity ($\sim 10^{-9}$
176 A m^2). The instrument was calibrated after every ten measurements by means
177 of a magnetically-stable rock specimen, as described by Kardel et al. (2011,
178 2012). The SIRM value was normalised to leaf dry weight ($10^{-5} \text{ A m}^2 \text{ kg}^{-1}$) and
179 each SIRM value was the average of two measurements on the same sample.

180

181 **2.4 Statistical analysis**

182 First, the normality of the data was tested for all considered variables by
183 Kolmogorov-Smirnov tests, and the data that did not meet normality were
184 normalised by logarithmic transformations.

185 Significant differences in elemental concentration between unwashed and washed
186 leaves were tested for each species and each polluted site (Boumehen and Babol)
187 separately by t-tests.

188 Significant differences in elemental concentration between washed leaves (*C.*
189 *lawsoniana* and *P. brutia* subsp. *eldarica*) for polluted sites and unwashed leaves
190 for the reference site (Babolsar) were tested separately by t-tests.

191 Differences in concentrations of unwashed leaves and also in leaf SIRM values
192 were tested between species and sites by using a two-way analysis of variance
193 (ANOVA) procedure and a Tukey-HSD test.

194 Pearson's correlation analysis was performed to evaluate the relationship between
195 SIRM values of unwashed leaves and all element concentrations of unwashed and
196 washed leaves.

197 A Principal Component Analysis (PCA) using varimax orthogonal rotation with
198 Kaiser normalization was applied on the elemental concentrations of unwashed
199 and washed leaves, and also on the elemental concentrations of the unwashed
200 leaves, to 'cluster' the sites and species, to estimate the potential common
201 sources and to reduce multicollinearity. Before the PCA, Kaiser-Meyer-Olkin
202 (KMO) measurement of sampling adequacy and Bartlett's test of sphericity were
203 used to test for data suitability. Statistical analysis was carried out using R
204 version 3.4.2 software (R Development Core Team 2017). The Fcatorextra and
205 ggplot2 Package (Wickham, 2009) and FactoMineR package (Le et al. 2008;
206 Husson et al. 2017) were used for PCA analysis and making PCA biplots.

207

208 **3 Results**

209 **3.1 Chemical composition of unwashed leaves vs washed leaves**

210 The concentrations of macro-elements in the unwashed leaves amounted up to 76
211 mg/g dry weight, and 239 µg/g dry weight for micro-elements (Table 1). In the
212 washed leaves, the macro-elements had concentrations of up to 62 mg/g and the
213 micro-elements up to 150 µg/g. The highest concentrations were found for Ca
214 and K. In the washed leaves of one or more species, the concentrations of Si, Cr
215 and Co were below the detection limit. The Pb concentrations were below
216 detection limit in all washed and unwashed samples.

217 The order of abundance of deposited macro-elements on the leaf surface, based
218 on the difference between the unwashed and washed leaves, was Ca > Si > Fe >
219 Al > Mg > Ti > Na > Mn, and for micro-elements was Zn > Cr > Ba > Ni > Cu > Sb >

220 Zr> V> Rb> Co. The unwashed leaves of *P. brutia* subsp. *eldarica* in the city of
 221 Boumehen had significantly higher concentrations of Al, Ca, Fe, Mg, Mn, Na, Si,
 222 Ti, Ba, Co, Cr, Cu, Ni, Rb, V, Zn, and Zr elements compared to its washed leaves
 223 ($p<0.05$) (Table 1). The unwashed leaves of *C. lawsoniana* from Babol city had
 224 significantly higher concentrations of Al, Fe, Ca, Mn, Si, Ti, Ba, Co, Cr, Cu, Ni, Rb,
 225 Sb, V, Zn and Zr elements compared to the washed leaves of this tree species
 226 ($p<0.05$) (Table 1). The unwashed leaves of *L. japonicum* in the city of Babol
 227 had, compared to the washed leaves, significantly higher concentrations of Al, Ca,
 228 Fe, Mg, Si, Ti, Co, Cu, Ni, V, and Zr elements ($p<0.05$) (Table 1). For Cl, K, P, S,
 229 As, Cd, Cs, Pb, Sn, and Sr elements, there was no significant difference between
 230 washed and unwashed leaves of any of these species. So for further analysis,
 231 these elements are not considered.

232 Moreover, the PCA biplot on all chemical elements of unwashed and washed
 233 leaves showed higher values of all considered elements that relate to the first axis
 234 for unwashed leaves compared to the washed leaves (Fig. 2). The elements that
 235 relate to the first axis were Al, Fe, Ca, Si, Ti, Co, Cr, Cu, Ni, Rb, V, Zn and Zr.

236

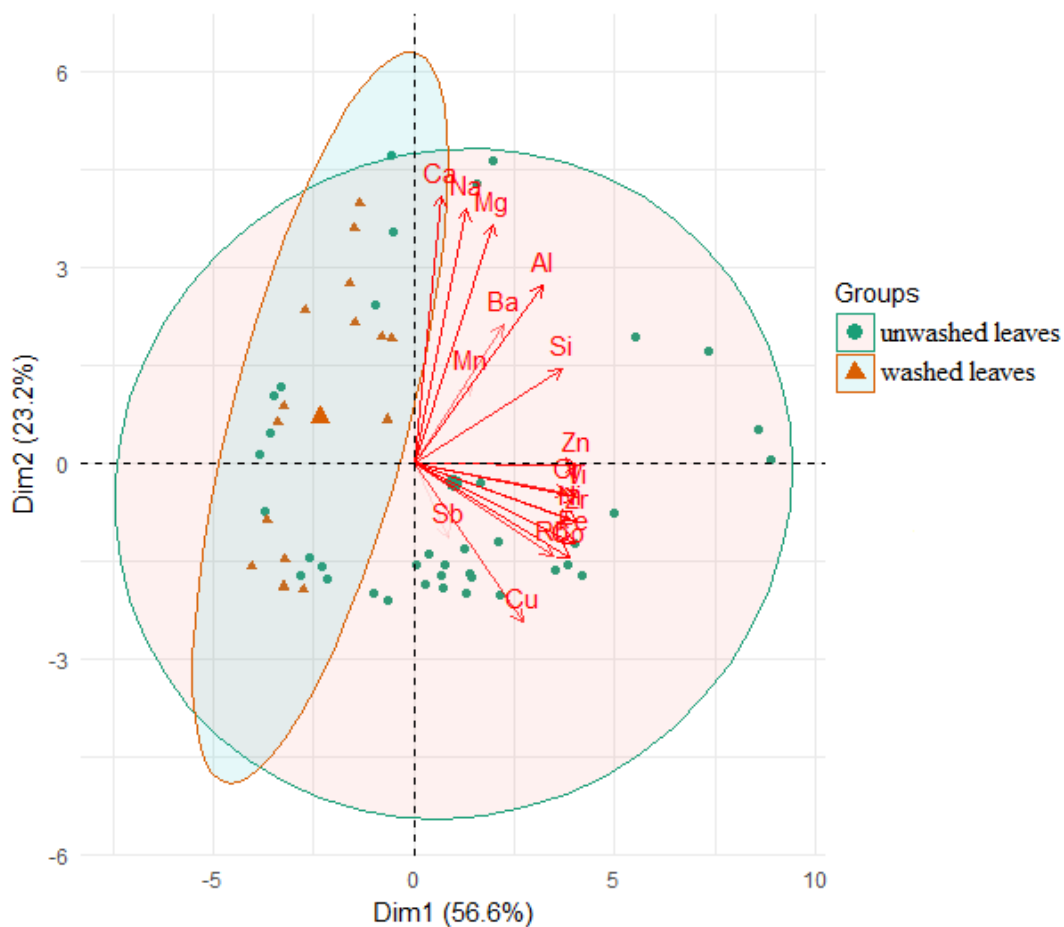
237 Table 1: Mean (\pm standard deviation) of elemental concentrations (in mg/g dry
 238 weight or $\mu\text{g/g}$ dry weight) in unwashed and washed leaves of *L. japonicum*, *C.*
 239 *lawsoniana* and *P. brutia* subsp. *eldarica* (Medw) Silba, from Boumehen and
 240 Babol. Significant differences ($p<0.05$) between unwashed and washed leaves are
 241 shown in bold.

	<i>P. brutia</i> subsp. <i>eldarica</i>		<i>C. lawsoniana</i>		<i>L. japonicum</i>	
	Unwashed	Washed	Unwashed	Washed	Unwashed	Washed
Macro-elements (mg/g)						
Al	3.28±0.71	1.70±0.18	5.48±1.04	4.21±0.38	5.01±0.62	3.91±0.44
Ca	23.75±3.63	15.02±2.09	52.74±5.90	41.06±4.21	76.33±9.89	62.27±9.95
Cl	3.45±1.77	3.50±1.30	2.11±0.95	2.52±0.97	1.70±0.32	1.92±0.59
Fe	6.21±1.52	2.31±0.33	8.90±1.96	4.15±0.62	3.13±0.53	1.41±0.04

K	26.34±4.06	26.80±1.88	40.55±7.36	32.68±8.27	16.98±2.84	19.97±3.02
Mg	1.93±0.39	1.10±0.18	3.67±0.82	3.32±0.46	3.90±0.47	3.15±0.42
Mn	0.19±0.02	0.14±0.02	0.14±0.03	0.11±0.01	0.23±0.05	0.29±0.11
Na	1.29±0.08	1.14±0.03	1.70±0.19	1.69±0.13	1.82±0.11	1.65±0.11
P	3.09±0.25	2.79±0.18	2.91±0.57	3.32±0.58	3.02±0.21	2.77±0.46
S	3.85±0.87	2.38±0.68	4.76±0.63	3.60±1.19	5.66±0.65	4.61±1.49
Si	2.36±2.41	DL	8.10±3.17	0.40±0.27	5.41±2.54	2.27±0.89
Ti	0.40±0.09	0.13±0.02	0.81±0.19	0.35±0.08	0.25±0.06	0.09±0.00
Micro-elements (µg/g)						
As	11.50±0.51	11.50±0.57	9.80±1.52	9.60±0.00	10.20±0.45	10.25±0.50
Ba	80.50±32.56	22.75±8.18	80.20±25.96	32.50±7.85	143.60±33.90	150.75±24.05
Cd	2.00±0.00	2.00±0.00	2.00±0.00	2.00±0.00	2.00±0.00	2.00±0.00
Co	2.10±0.64	0.75±0.50	3.20±0.84	1.00±0.00	0.80±0.45	DL
Cr	53.10±20.51	DL	165.80±45.80	21.25±13.15	17.40±19.02	DL
Cs	5.15±0.59	4.50±0.57	4.00±1.22	3.75±0.50	4.60±0.55	4.75±0.50
Cu	61.25±21.02	19.25±6.94	38.40±14.65	15.50±4.43	8.60±3.78	1.50±1.91
Ni	33.55±7.36	12.00±2.16	64.00±15.05	19.75±4.11	18.40±5.18	7.50±1.00
Pb	DL	DL	DL	DL	DL	DL
Rb	4.85±1.66	1.25±0.50	5.60±1.52	1.00±0.00	2.20±0.84	1.75±1.26
Sb	40.05±11.56	29.75±13.32	19.40±11.57	3.75±2.63	36.80±8.90	35.75±8.42
Sn	1.10±0.31	1.00±0.00	1.20±0.45	1.25±0.50	1.00±0.00	1.00±0.00
Sr	93.65±13.73	81.50±14.05	161.80±23.40	168.75±47.38	147.2±22.55	144.50±38.56
V	10.00±3.01	2.00±0.82	17.20±3.42	8.00±0.00	6.00±1.58	3.25±0.50
Zn	116.05±31.70	37.25±14.59	239.80±79.05	69.00±17.00	85.40±53.40	56.25±49.54
Zr	8.75±2.57	2.50±0.58	16.00±4.06	4.75±1.71	4.60±1.82	2.00±0.00

242 DL: below detection limit

243



244

245 Fig. 2: PCA biplot for elements of unwashed and washed leaves in Boumehen and
 246 Babol sites.

247

248 **3.2 Chemical compositions of roadside tree leaves in different sites**

249 The statistical results for unwashed leaves revealed that species had a significant
 250 effect on the accumulation of the chemical elements. The unwashed leaves of *C.*
 251 *lawsoniana* had significantly ($p < 0.05$) higher concentrations of Fe, Ti, Co, Cr, Cu,
 252 Ni, Rb, Zn, Zr, V and lower concentrations of Ca, Mn, Ba, Sb compared to
 253 unwashed leaves of *L. japonicum* in the same site (Babol; Table 2). The unwashed
 254 leaves of *C. lawsoniana* had significantly higher concentrations of Al, Ca, Mg, Na
 255 and lower concentrations of Fe, Ti, Cr, Ni, Sb compared to unwashed leaves of *P.*
 256 *brutia* subsp. *eldarica* in the same site (Babolsar; $p < 0.05$; Table 2).

257 Moreover, the statistical results revealed that chemical compositions of unwashed
 258 leaves and their concentration highly depended on site ($p < 0.05$). The Boumehen
 259 and Babol sites had significantly higher concentration of all considered elements
 260 compared to the reference site in Babolsar, except for Ca, Rb and Sb in
 261 Boumehen, and Na in Babol, for the same species ($p < 0.05$) (Table 2).

262 The statistical results showed there was no significant difference between
 263 elements for washed leaves of *P. brutia* subsp. *eldarica* in Boumehen site
 264 compared to unwashed leaves of this species in the reference site at Babolsar
 265 except for Al, while the concentration of Al, Fe, Ti, Mn, Ni and Zn elements for
 266 washed leaves of *C. lawsoniana* were significantly ($p < 0.05$) higher in Babol site
 267 compared to unwashed leaves of this species in the reference site Babolsar (data
 268 are shown in Table 1 and 2).

269

270 Table 2: Chemical composition (mean±standard deviation) of unwashed leaves of
 271 *L. japonicum*, *C. lawsoniana* and *P. brutia* subsp. *eldarica* from different sites.
 272 Significant differences between sites for each species at $p = 0.001$ '***', $p = 0.01$
 273 '**', $p = 0.05$ '*'. Significant differences between co-located species are shown in
 274 bold.

	<i>P. brutia</i> subsp. <i>eldarica</i>		<i>C. lawsoniana</i>		<i>L. japonicum</i>
	Boumehen	Babolsar	Babol	Babolsar	Babol
Macro-elements (mg/g)					
Al	3.28±0.72***	2.06±0.12	5.48±1.04**	2.76±0.25	5.01±0.62
Ca	23.74±3.63	26.17±7.42	52.74±5.91*	39.81±7.26	76.33±9.89
Fe	6.21±1.52***	2.36±0.40	8.89±1.96***	1.53±0.19	3.13±0.52
Mg	1.93±0.39**	1.47±0.14	3.67±0.82*	2.34±0.27	3.90±0.47
Mn	0.19±0.02**	0.08±0.03	0.14±0.03**	0.06±0.02	0.23±0.05
Na	1.29±0.08*	1.21±0.05	1.70±0.19	1.50±0.08	1.82±0.11
Si	2.36±2.41***	DL	8.10±3.18**	DL	5.41±2.54
Ti	0.40±0.09***	0.18±0.04	0.81±0.19**	0.10±0.02	0.25±0.06
Micro-elements (µg/g)					

Ba	80.50±32.56***	12.75±4.11	80.20±25.96***	17.75±3.09	143.60±33.90
Co	2.10±0.64**	0.75±0.50	3.20±0.84**	DL	0.80±0.45
Cr	53.10±20.51***	17.00±9.09	165.80±45.81**	1.00±2.00	17.40±19.02
Cu	61.25±21.02***	12.50±9.25	38.40±14.65**	10.75±3.50	8.60±3.78
Ni	33.55±7.36***	18.75±3.50	64.00±15.05**	12.00±2.16	18.40±5.18
Rb	4.85±1.66	1.75±2.22	5.60±1.52***	0.75±0.96	2.20±0.84
Sb	40.05±11.56	34.50±7.23	19.40±11.57*	3.25±3.95	36.80±8.90
V	10.00±3.01***	4.75±1.90	17.20±3.42***	2.25±1.26	6.00±1.58
Zn	116.05±31.70***	1.75±3.50	239.80±79.06**	DL	85.40±53.46
Zr	8.75±2.57***	3.50±1.00	16.00±4.06**	2.00±0.00	4.60±1.82

275

276 3.3 Leaf SIRM of roadside tree leaves in different sites

277 The leaf SIRM values of all species and sites varied between 0.03 and 893.49 x
 278 10^{-5} A m² kg⁻¹ (Table 3). Both *C. lawsoniana* and *P. brutia* had significantly
 279 ($p < 0.0001$) higher leaf SIRM values in Babol and Boumehen than in reference
 280 site Babolsar.

281 In Babol, leaves of *C. lawsoniana* had significantly higher SIRM values compared
 282 to *L. japonicum* ($p < 0.05$), but there was no significant difference between leaf
 283 SIRM values of co-located species (*C. lawsoniana* and *P. brutia*) in Babolsar.

284

285 Table 3: Mean (\pm SD) SIRM ($\times 10^{-5}$ A m² kg⁻¹) for unwashed leaves of *L.*
 286 *japonicum*, *C. lawsoniana* and *P. brutia* subsp. *eldarica* from Boumehen, Babol
 287 and the reference site Babolsar. Significant differences ($p < 0.05$) are shown with
 288 different letters.

Species	Site	SIRM $\times 10^{-5}$ A m ² kg ⁻¹
<i>P. brutia</i> subsp. <i>eldarica</i>	Boumehen	696.89±83.66 ^c
	Babolsar	0.06±0.01 ^a
<i>C. lawsoniana</i>	Babol	783.97±92.73 ^d

	Babolsar	0.05±0.02 ^a
<i>L. japonicum</i>	Babol	629.69±44.09 ^b

289

290

291 **3.4 Correlation between leaf SIRM and deposited elements on leaf**
 292 **surface**

293 There was a strong, positive and significant correlation between Fe, Ti, Co, Cr,
 294 Cu, Ni, Rb, V, Zn, and Zr elements and between Al, Ca, Mg, Na, Si, and Ba
 295 elements in unwashed roadside leaves (Table 4a), and the same correlations
 296 persisted for washed leaves except for Rb (Table 4b). In unwashed leaves, a
 297 positive and significant correlation were observed between Si, and Al, Ca, Fe, Mg,
 298 Na, Ba, Co, Cr, Ni, Rb, V, Zn, Zr, SIRM, while in washed leaves, a significant and
 299 positive correlation persisted between Si and Al, Ca, Mg, Mn, Na, Ba, Zn.
 300 Moreover, a positive and significant correlation was observed between leaf SIRM
 301 and Al, Fe, Mn, Si, Ti, Ba, Co, Cu, V, Zn, Zr elements of unwashed roadside
 302 leaves, while for washed leaves were observed a positive and significant
 303 correlation between SIRM and Fe, Co, Ni, Zn elements (Table 4).

304

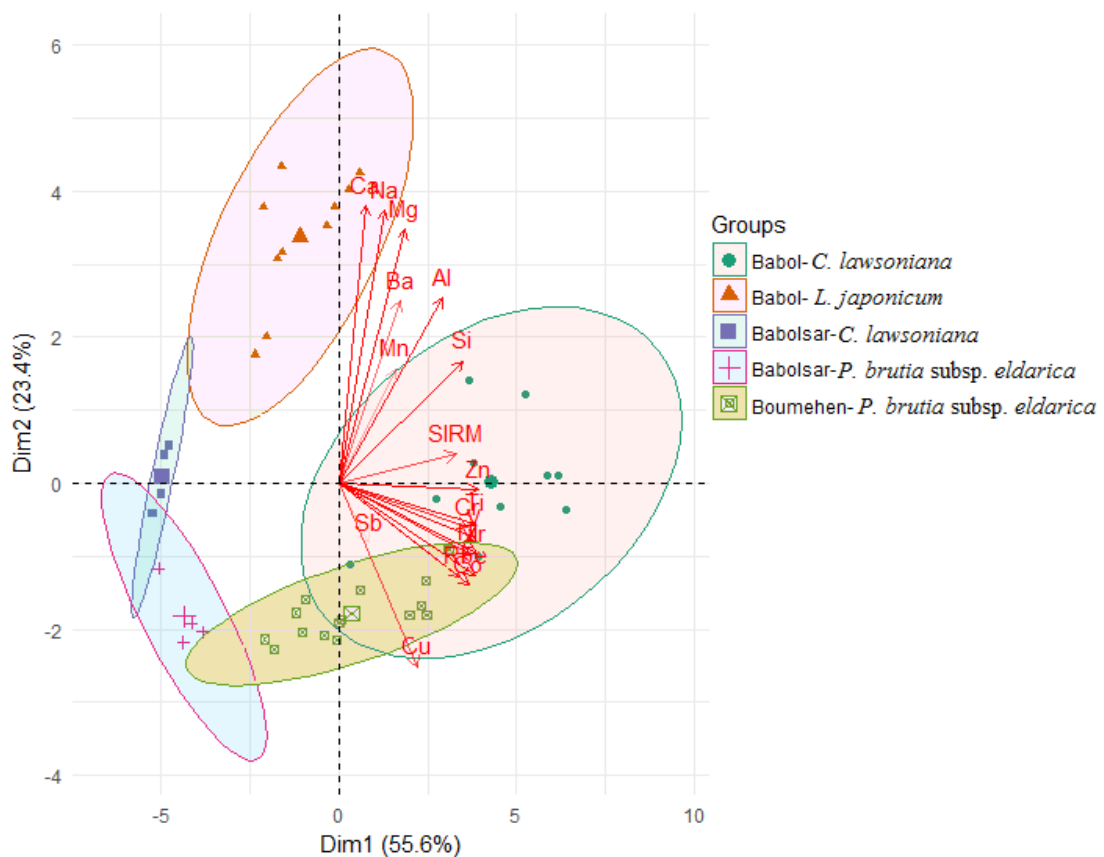
305 **3.5 PCA analysis on elemental concentrations and SIRM of unwashed**
 306 **leaves**

307 First, the result of Bartlett test was significant ($p < 0.05$) and the criterion of
 308 Kaiser Meyer-Olkin (KMO) obtained 0.79 which confirms the suitability of the data
 309 for further analysis using PCA on the unwashed leaves. The components with
 310 eigenvalues greater than one were considered for varimax rotation to obtain the
 311 final matrix. Also factor loadings greater than 0.50 were considered significant.
 312 The first component of the PCA for unwashed leaves explained 55.58 % of the
 313 total variance and showed strong positive loadings for Fe, Ti, Co, Cr, Ni, V, Zn
 314 and Zr, and moderate positive loadings for Al, Si, Cu, Rb and SIRM. The second

315 factor explained 23.37 % of the total variance with high positive loadings of Ca,
316 Mg, Na and Ba. The third factor explained 12.02 % of total variance with positive
317 loadings Mn and Sb.

318 The biplot for unwashed leaves (Fig. 3) revealed that the first component
319 differentiated the roadside leaves (Boumehen and Babol) from the leaves at the
320 reference site (Babolsar), and it also differentiated between *C. Lawsoniana* and *L.*
321 *Japonicum* in Babol. The second component differentiated between species (Fig.
322 3). Moreover, the biplot showed that SIRM differentiated between sites and
323 species in the same way as the derived chemical elements from traffic and soil
324 dust (Fig. 3).

325



326

327 Fig. 3: PCA biplot for elements and SIRM value of unwashed leaves in different
328 sites and species

329

330 **4 Discussion**

331 **4.1 Chemical compositions of leaves**

332 When using heavy metal concentration of plant leaves for biomonitoring of
333 atmospheric pollution, it is often difficult to distinguish the atmosphere
334 (deposition) from the soil (root uptake) as sources of the observed metal
335 concentration in the leaf sample. Leaf washing is a solution to overcome this
336 problem (Al-Alawi and Mandiwana 2007). The differences in element
337 concentrations between washed and unwashed leaves revealed that deposited
338 particles on the leaf surface in the studied sites may contain various elements
339 including Mg, Al, Si, P, S, Ca, Ti, V, Cr, Mn, Fe, Co, Ni, Cu, Zn, Rb, Zr, Sb and Ba.
340 Our results confirm that the atmospheric deposited particles contain various
341 metals which are in line with finding of other studies. Al-Alawi and Mandiwana
342 (2007) used the washing technique for the leaves of *Pinus halepensis* L. and
343 reported that the unwashed leaves had significantly higher content of Pb, Cu, Zn
344 and Cd compared to the washed leaves, confirming that atmospheric deposition
345 contains metals in different sites of Amman City, Jordan. Oliva and Espinosa
346 (2007) investigated metal concentration in soil, plant leaves and the atmosphere,
347 and they reported no clear correlation between soil and plant leaves. They stated
348 that their results indicate that the considered metal concentration in plant leaves
349 mostly related to atmospheric metal concentration.

350 The results of leaf washing in our study showed that there was no significant
351 difference between washed and unwashed leaves for Cl, K, P, S, As, Cd, Cs, Pb,
352 Sn, and Sr elements. These results indicate that tree leaves may not be a suitable
353 biomonitor for these elements or request other methodology for the detection of
354 these elements. The possible reasons are: a) these elements are just not present
355 in the atmosphere at high concentrations and so deposition is low (e.g. P, Pb, Sn
356 and Cd) or b) some of these elements are present in such high concentration in
357 the leaf, that low deposition of these elements (such as K) on the leaves are not
358 sufficient to considerably increase the total element concentrations of the

359 unwashed leaves, and c) elements such as K leach in huge amounts from the
360 leaves when immersed in the water, which could have increased the K
361 concentration in the washing solution very fast and therefore the remaining drops
362 of washing solution on the leaves may have still contributed a lot to the
363 concentrations of the washed leaves. Further research is needed on soil,
364 atmospheric dust and tree leaves to identify the source of these elements (Cl, K,
365 P, S, As, Cd, Cs, Pb, Sn, and Sr) in the atmospheric deposited particles. Our
366 results on concentrations of water soluble ions in the washing solution in the
367 same samples revealed significantly high concentration of SO_4^{2-} in polluted sites
368 (Boumehen and Babol) compared to Babolsar and coming from atmosphere (data
369 not published), while in the present study there was no significant difference in S
370 concentration between unwashed/washed leaves in polluted sites, probably
371 because these SO_4^{2-} ions in the washing solution were coming from leaching out
372 of the interior of the leaf. However, its concentration was significantly higher in
373 Boumehen and Babol sites compared to Babolsar site (data not shown).

374 Serbula et al. (2012) assessed heavy metal concentrations in different organs of
375 *Robinia pseudoacacia* L. and topsoil in an industrial area in Bor (Eastern Serbia).
376 They reported that *R. pseudoacacia* is not a suitable indicator of air and soil As
377 pollution as the As concentration was high in air and topsoil but not in the plant
378 organs. Moreover, in agreement with our results, Norouzi et al. (2015)
379 investigated the concentrations of Cu, Fe, Mn, Ni, Pb, and Zn elements of
380 *Platanus orientalis* L. tree leaves in Isfahan, Iran. They reported *P. orientalis* tree
381 leaves as a suitable bioindicator for atmospheric pollution for all considered heavy
382 metals except Pb.

383 In comparison with other studies, our results for roadside trees showed elevated
384 values for most considered element concentrations, and leaf SIRM. For instance,
385 all considered tree species in Babol and Boumehen sites in our study showed a
386 higher value for Ba, Cd, Cr, Cu, Fe, and Zn elements in leaves comparing to

387 leaves of *Quercus ilex* L. (also an evergreen species) in Florence, Italy, except for
388 Mn and Pb (Ugolini et al. 2013). Moreover, the concentration of Al, Ba, Cd, Co,
389 Cr, Cu, Fe, Ni, Sr, V, and Zn elements and the SIRM value for all considered
390 species on our roadside sites (Babol and Boumehen) were much higher compared
391 to two moss species (*Sphagnum girgensohnii* and *Hypnum cupressiforme*)
392 exposed over the city of Belgrade for short time exposure (during summer)
393 except for Pb (Vuković et al. 2015), as Pb was not detected in our study.

394

395 **4.2 Effect of sites and species on leaf elements and SIRM**

396 Our results showed a significant effect of site and species on the leaf elemental
397 concentrations and the leaf SIRM, and the effect of site is more pronounced than
398 the effect of species in the variation of chemical composition.

399 The results of chemical composition of the leaf showed that the concentration of
400 many heavy metals is different for the different sites and species, and the PCA
401 biplot confirms these results accordingly (Fig 3). For instance, the high
402 concentration of Cu and Sb elements for tree leaves of *Pine* sp. in the Boumehen
403 compared to the Babolsar indicates that the source of pollution in Boumehen site
404 is different from that in the Babolsar site or the same source but further away
405 from the sampled trees. As the differences between the roadside leaves and the
406 reference site leaves mainly occurred along the first axis of the PCA, the roadside
407 and reference sites thus mainly differed in element concentrations that showed
408 high factor loadings with this axis, i.e. Fe, Ti, Co, Cr, Ni, V, Zn, Zr and to a lesser
409 extent Al, Si, Cu, Rb and SIRM. Likewise, differences in elemental concentrations
410 between species at the same site were most pronounced in the concentrations of
411 elements that had high factor loadings with the second axis of the PCA, i.e. Ca,
412 Na, Mg, Ba and Cu. We can conclude that the presence of road traffic mainly
413 leads to an increase in metals such as Fe, Ti, Co, Cr, Ni, V, Zn and Zr, and that
414 differences between species are most pronounced in elements like Ca, Na and Mg
415 which are mainly deposited as larger particles.

416 Significant differences in the metal concentration of leaves for the co-located
417 species in Babol are related to the plant structure and the leaf surface
418 characteristics. Our results showed that *C. lawsoniana* with micro-roughness leaf
419 surface had significantly higher metal concentrations and SIRM values compared
420 to *L. japonicum* with smooth and waxy leaf surface in Babol. These results
421 confirm a higher ability of *C. lawsoniana* to accumulate atmospheric particles in
422 comparison with *L. Japonicum*. In the Babolsar site, *P. brutia* subsp. *eldarica* had
423 significantly higher concentration of metals with atmospheric sources and lower
424 soil dust and sea salt sources compared to *C. lawsoniana*. The reason could be
425 that the trees were at a close distance from the road for *P. brutia* subsp. *eldarica*,
426 while the *C. lawsoniana* trees were closer to the beach. The effect of distance of
427 tree leaves from the roads has been reported by other studies too (e.g., Matzka
428 and Maher 1999; Szönyi et al. 2008; Kardel et al. 2012).

429 Our results are in agreement with those of Moreno et al. (2003) who first pointed
430 out the effect of species and Freer-Smith et al. (2005), who reported an effect of
431 species and site on atmospheric-particle accumulation on tree leaves. An effect of
432 species on atmospheric-particle accumulation has been reported by other studies
433 too (Amato-Lourenco et al. 2016; Jamil et al. 2009; Sæbø et al. 2012; Song et al.
434 2015).

435 Wang et al. (2006) observed plant with a large micro-roughness leaf surface
436 accumulates a large amount of dust particles. Wang et al. (2015) observed
437 epicuticular wax ultrastructures significantly contributing to particle accumulation
438 on leaf surfaces. Simon et al. (2014) reported no significant difference in the
439 amount of deposited particles along an urbanization gradient due to metrological
440 conditions and topography, but deposited particles depended on species
441 characteristics (e.g. trichomes density on leaf surface). A study by Sæbø et al.
442 (2012) on 47 tree and shrub species identified hair density, wax quantity and
443 specific leaf area (leaf area to leaf dry weight ratio) as leaf surface traits that
444 affect PM accumulation. Our observations of the differences in species are in line

445 with the relationship of higher SLA with higher particle deposition as observed by
446 Sæbø et al. (2012) and Wang et al. (2006). Next to that, also leaf longevity and
447 related time exposure play an important role in the species effect on
448 accumulation of particles on leaves (Lehndorff et al. 2006). An effect of leaf
449 surface characteristics and exposure time on magnetic particles accumulation has
450 been reported by other studies too (Hofman et al. 2014; Kardel et al. 2011;
451 Mitchell et al. 2010).

452

453 **4.3 Deposited metals: sources and their relation to leaf SIRM**

454 The link between Co, Cr, Cu, Fe, Ni, Ti, V, Zn, Zr metals and motorised road
455 traffic is known from literature and these metals occur in the unwashed leaves, so
456 the presence of these metals in the leaves of Boumeham and Babol strongly
457 indicate that road traffic is an important contributor to the atmospheric pollution
458 in these sites. In the first component of the PCA for unwashed leaves, moderate
459 loadings of Al, Cu and Si indicate that these elements co-occurred with elements
460 derived from traffic and have other sources too. Moreover, the good correlation of
461 Al, Cu, Rb and Si with leaf SIRM for unwashed leaves indicates that leaf SIRM is a
462 suitable indicator of these elements. The second component has high positive
463 loadings of Ca, Mg, Na, and Ba, which are considered as crustal and sea salt
464 elements, indicating that soil dust and sea salt could be another important source
465 of deposited particles. Moreover, elements from sea salt and marine spray were
466 not significantly correlated with leaf SIRM values since they are diamagnetic
467 (Sagnotti et al. 2006), so that leaf SIRM could not serve as a proxy for these
468 elements. The third component has positive loadings of Mn and Sb, indicating
469 that these elements have another source comparing to the two first components,
470 which could be industrial activities. Moreover, leaf SIRM also can be used for
471 monitoring of industrial activities when the activities produce magnetic particles
472 (Hanesch et al. 2003; Schädlich et al. 1995).

473

474 The leaf SIRM correlated significantly and positively with the elemental
475 concentrations of unwashed leaves except for Na, Mg, Ca, Sb and Rb, while most
476 relationships weakened or even disappeared when the leaves were washed. These
477 findings suggest that leaf SIRM is mainly related to the metals deposited on the
478 leaf surface and are in agreement with other studies. For example, Lu et al. 2008
479 report a high correlation between magnetic properties and heavy metal content
480 (mainly Fe, Mn, and Cu) of deposited dust on leaf surface in China. Also, Wang
481 (2016) found high correlation of Cu and Ba with SIRM in street dust in Xuzhou,
482 China. Castanheiro et al. (2016) were able to relate leaf SIRM with the Fe, Zn and
483 Pb content of particles deposited on *Hedera* sp. leaves in an urban environment in
484 Belgium, but also significant relationships with the metals Cd and Mn were found.
485 Several former studies introduced leaf SIRM as good proxy of atmospheric
486 pollution, traffic-derived or industry-derived particles and dust pollution (Hansard
487 et al. 2011; Hofman et al. 2014; Kardel et al. 2012; Maher et al. 2008), which is
488 confirmed in this study.

489

490 **5 Conclusions**

491 Our results revealed that the leaves of roadside trees are able to accumulate on
492 their surface various elements including Al, Fe, Ca, Mg, Mn, P, S, Si, Ti, Ba, Co,
493 Cr, Cu, Ni, Sb, Rb, V, Zn, Zr. Thus, it can be concluded that leaves can be a
494 suitable monitor for these deposited elements. While leaf surface may not be a
495 good indicator for deposition of some metals such as K, Cl, P, S, As, Cd, Cs, Pb,
496 Sn and Sr elements.

497 From the chemical composition and SIRM value of the three considered species, it
498 can be concluded that *Chamaecyparis lawsoniana* has the highest potential to
499 accumulate atmospheric particles for monitoring purposes in Iran as well as for
500 other places where this species is available.

501 Fe, Ti, Co, Cr, Ni, Rb, V, Zn, and Zr elements were abundantly observed on
502 leaves in two sites and these elements are known in literature to originate from

503 motorized road traffic, which indicates the main source of pollution in Babol and
504 Boumehen is nearby road traffic rather than other, more distant sources. The
505 most abundant traffic-derived metals showed to be strongly correlated with leaf
506 SIRM. The leaf SIRM and the concentration of traffic-derived metals clearly
507 depend on sites and species with the effect of site being much stronger than the
508 effect of species. Overall, it can be concluded that leaf SIRM is a good proxy for
509 anthropogenic particles e.g. derived from traffic and/or industrial activity, but is
510 not a suitable indicator for sea salt elements.

511 This study, by indicating suitable trees and appropriate analytical methods for air
512 pollution monitoring, provides valuable information for further urban
513 managements and planning, and biomonitoring of atmospheric PM pollution in
514 Iran and anywhere else where the studied trees occur.

515

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520

521

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698

Table 4: Correlation matrix for the SIRM of unwashed leaves and the concentrations of macro and micro elements ($\mu\text{g/g}$) of a) unwashed leaves and b) washed leaves. Significant correlations are shown in bold ($p < 0.05$).

a)

	Macro elements							Micro elements										Magnetic parameter
	Ca	Fe	Mg	Mn	Na	Si	Ti	Ba	Co	Cr	Cu	Ni	Sb	Rb	V	Zn	Zr	SIRM
Al	0.70	0.57	0.93	0.45	0.85	0.94	0.69	0.74	0.53	0.63	0.09	0.60	0.05	0.46	0.66	0.74	0.64	0.56
Ca		-0.11	0.89	0.26	0.93	0.53	0.10	0.50	-0.15	0.11	-0.49	0.02	-0.22	-0.19	0.03	0.17	0.01	0.08
Fe			0.25	0.41	0.09	0.68	0.94	0.46	0.97	0.84	0.75	0.89	0.28	0.88	0.95	0.91	0.97	0.76
Mg				0.31	0.98	0.80	0.44	0.63	0.21	0.42	-0.21	0.35	-0.13	0.15	0.38	0.49	0.35	0.32
Mn					0.22	0.43	0.28	0.80	0.35	0.13	0.44	0.20	0.56	0.43	0.33	0.48	0.32	0.76
Na						0.69	0.31	0.54	0.05	0.30	-0.34	0.21	-0.23	0.00	0.22	0.35	0.20	0.18
Si							0.79	0.71	0.65	0.71	0.20	0.70	0.16	0.59	0.76	0.83	0.75	0.55
Ti								0.40	0.92	0.95	0.49	0.96	0.12	0.76	0.96	0.94	0.98	0.67
Ba									0.42	0.23	0.32	0.29	0.46	0.52	0.44	0.55	0.44	0.64
Co										0.81	0.72	0.87	0.32	0.88	0.93	0.88	0.95	0.73
Cr											0.34	0.99	-0.01	0.63	0.91	0.86	0.89	0.55
Cu												0.45	0.42	0.75	0.55	0.53	0.61	0.63
Ni													0.09	0.72	0.95	0.89	0.93	0.60
Sb														0.45	0.21	0.25	0.21	0.39
Rb															0.82	0.77	0.85	0.40
V																0.89	0.95	0.68
Zn																	0.95	0.77
Zr																		0.77

b)

	Macro elements							Micro elements										Magnetic parameter
	Ca	Fe	Mg	Mn	Na	Si	Ti	Ba	Co	Cr	Cu	Ni	Sb	Rb	V	Zn	Zr	SIRM
Al	0.91	0.31	0.99	0.26	0.98	0.57	0.40	0.52	0.08	0.51	-0.35	0.33	-0.43	-0.21	0.66	0.50	0.30	0.29
Ca		-0.01	0.93	0.42	0.92	0.73	0.00	0.74	-0.24	0.23	-0.63	-0.01	-0.15	-0.03	0.42	0.34	0.04	0.08
Fe			0.22	-0.32	0.19	-0.28	0.98	-0.40	0.83	0.81	0.54	0.92	-0.76	-0.36	0.85	0.50	0.91	0.54
Mg				0.22	0.99	0.54	0.31	0.53	-0.00	0.44	-0.39	0.25	-0.37	-0.16	0.60	0.40	0.20	0.18
Mn					0.20	0.84	-0.33	0.71	-0.25	-0.23	-0.53	-0.35	0.18	0.01	-0.17	0.61	-0.30	0.35
Na						0.51	0.28	0.51	-0.02	0.43	-0.39	0.24	-0.36	-0.15	0.57	0.37	0.17	0.16
Si							-0.22	0.91	-0.38	-0.09	-0.64	-0.28	0.15	0.07	0.00	0.51	-0.17	0.24
Ti								-0.34	0.73	0.84	0.43	0.88	-0.77	-0.32	0.89	0.47	0.93	0.43
Ba									-0.44	-0.21	-0.68	-0.41	0.29	0.02	-0.08	0.25	-0.28	0.16
Co											0.56	0.70	0.80	-0.43	0.60	0.43	0.63	0.63
Cr												0.31	0.91	-0.70	0.81	0.48	0.84	0.41
Cu													0.15	-0.02	0.12	0.07	0.47	0.25
Ni														-0.70	0.78	0.47	0.85	0.52
Sb														0.69	-0.73	0.48	-0.62	-0.51
Rb															-0.33	0.48	-0.19	-0.56
V																0.48	0.75	0.43
Zn																	0.47	0.70
Zr																		0.38